Final Report

to the

Southeast Area Monitoring and Assessment Program - Caribbean

United States Virgin Islands Queen Conch Stock Assessment

Executive Summary

Conch stocks have been periodically surveyed in the U.S. Virgin Islands. Previous conch scooter transect surveys were completed in 1981, 1985, 1990, and 1996. In 2001, a follow-up queen conch scooter transect survey study was completed. The objective of this study was to determine the status of conch stocks, assess the effectiveness of management regulations, and make management suggestions.

During this study, conch densities, particularly adults, were higher on scooter transects around St. Thomas compared to St. John. Overall conch densities around St. Croix were higher than St. Thomas and St. John. Mean juvenile conch densities for all transects common to all sampling years ranged from 1.0 juvenile conch/ha on St. Thomas to 72.3 juvenile conch/ha on St. Croix. St. John had 7.5 juvenile conch/ha. Mean adult conch density for all common transects was similar on St. Thomas (24.2 adult conch/ha) and St. Croix (27.4 adult conch/ha), but much lower on St. John (7.2 adult conch/ha).

Both adult and juvenile conchs were more abundant on algal plains and seagrass beds than other benthic habitats. Maximum adult conch densities were found in the 13 to 18 m depth range for common transects around St. Thomas and St. John. In St. Croix, the highest mean adult densities were found in the 19 to 24 m depth range. In general, there was a decrease in juvenile conch with increasing water depth.

Results of this study were compared with previous studies. Between 1996 and 2001, there was a substantial decrease in overall queen conch densities for common transects in St. Thomas. Between 1981 and 2001, there was also a substantial decrease in overall queen conch densities for the 9 common transects common to all survey years for St. John. This decline occurred despite current bag limits, minimum size limits, and seasonal closure. Over the past two decades, St. Croix had higher overall queen conch densities than either St. Thomas or St. John. Compared to previous years, juvenile conch densities in 2001 were much lower on St. John and St. Thomas. Low juvenile conch densities may reflect the lack of successful recruitment, patchy distribution of the species, and/or more importantly intensive fishing pressure.

It is recommended that surveys in deeper water $(>=24m)$ be conducted in the future. These depths have not previously been surveyed. In order to obtain a more complete picture of the status of conch stocks, these depths need to be surveyed. Other management recommendations include more public education, additional resource surveys and restoration efforts, and stricter enforcement.

INTRODUCTION

The queen conch, *Strombus gigas,* is a large gastropod that has great economic significance throughout the Caribbean and adjacent regions (Appeldoorn 1994a). According to Appeldoorn (1994a), from 1988 to 1991 (four years) landings of queen conch throughout the Caribbean region were on the order of 4,000 metric tons (mt) of conch meat with a potential value of US \$ 40,000,000. The high market value for conch makes the resource an important source of foreign exchange, either through domestic sale within the tourist industry or through export. Accessibility, ease of harvest, and high demand has resulted in generally overfished stocks around the Caribbean.

The populations of queen conch around St. Thomas, St. John, and St. Croix, U.S. Virgin Islands were recognized as seriously depleted by the late 1970's (Wood and Olsen 1983). By the early 1980's queen conch harvest had exceeded sustainable yield around St. Thomas, St. John, and St. Croix, U.S. Virgin Islands (USVI) (Wood and Olsen 1983). Declines in landings were also reported by the CFMC (1988).

In response to severe overfishing of conch in St. Thomas/St. John District, the USVI government established a 5-year moratorium on the harvest of queen conch on St. Thomas/St. John beginning February 1988 (Friedlander 1997). Based on the results of 5 year closures conducted in Cuba and Bonaire (Munoz et al 1987), a 5-year closure was considered an appropriate time period to enhance stock size in the Virgin Islands. Unfortunately, there were no comprehensive surveys of queen conch populations at the end of the moratorium to determine if the closure was successful in enhancing the population size of queen conch. Also, after the USVI moratorium was lifted in 1992, there were no restrictions on conch harvest for two years. During the period of no restrictions, heavy fishing pressure was reported on conch stocks throughout the northern USVI (Friedlander 1997).

On 12 July 1994 USVI regulations on conch were approved establishing a closed season from July 1 to September 30 each year, a limit on the number of conch landed, a size limit, and a requirement to land conch whole and in the shell (VIRR 1994, see also CFMC 2001). The requirement to land conch in the shell created extensive controversy on St. Croix as conch shells piled up at landing sites and on beaches. However, this was considered the only way of effectively enforcing size regulations. Because of the controversy and lack of enforcement, regulations are often ignored, particularly on St. Croix where there is a fairly substantial conch fishery in comparison with St. Thomas/St. John.

The U.S. Virgin Islands landed approximately 15 mt (30,000 kg) of conch in the year 1990 (Appeldoorn 1994a). USVI commercial catch reports indicate that there is only a small commercial fishery for queen conch in St. Thomas and St. John ranging from 225.5 to 3,138 kg between the July 1993 to June 1994 and July 2000 to June 2001 fishing years (DFW data files). In St. Croix, reports indicate that conch landings doubled from 35,572 lb in July 1993 to June 1994 to 72,902 lb in July 1997 to June 1998 (Table 1).

The United States Virgin Islands encompasses three major islands: St. Thomas, St. John, and St. Croix, as well as several offshore keys (see Figure 1). The territory as a whole consists of a

variety of habitats including fringing and patch reefs, mangrove lagoons, seagrass beds, carbonated pavement, sand plains, and algal plains (Dammann and Nellis 1992). Available conch habitat in the USVI is mainly limited to waters of the surrounding shallow insular shelf platform (Appeldoorn 1987). The approximate size of this habitat was established to be 34,300 ha for St. Croix and 162,925 ha for St. Thomas and St. John (CFMC 1999). Within many of the surrounding bays, seagrass beds and algal plains provide habitat to both adult and juvenile queen conch.

Over the last 20 years, the Division of Fish and Wildlife conducted visual surveys in order to monitor the status of the U.S. Virgin Islands conch stock. Repeated sampling at selected sites can indicate trends in stock condition over time. Estimates of conch abundance around St. Thomas, St. John, and St. Croix were previously conducted in 1981 (Wood and Olsen 1983), 1985 (St. John only; Boulon 1987), 1990 (Friedlander et al 1994), and 1996 (Friedlander 1997).

OBJECTIVES

The objective of this project is: (1) to provide baseline information on the population status of queen conch in a variety of critical marine habitats in USVI; (2) to provide time series information on changes in the USVI queen conch population; (3) to monitor queen conch populations in marine reserve areas and other protected habitats, and (4) to collect, analyze, manage and disseminate fisheries independent data on the queen conch resources in the USVI.

METHODS

Study Sites

In 1981, an initial 22 transects located around the islands of St. Thomas and St. John were surveyed to determine the density and size distribution of queen conch (St. Thomas=10 transects, St. John=12 transects) (Wood and Olsen 1983). An additional site (Saba Island) was added to the St. Thomas transect list in 2001 because juvenile conch appeared to be abundant at this site (based on DFW staff observations).For 2001, a total of 11 sites were surveyed in St. Thomas, and 12 sites were surveyed in St. John (Figure 1 and Table 2).

There were also 22 original sites around the island of St. Croix in 1981 (Wood and Olsen 1983). In 2001, only the first 16 of the original 22 transects were resurveyed because of limited study time (Figure 1 and Table 2).

Scooter transects

Queen conch abundance and density estimates were derived from visual surveys conducted along transects by two scuba divers using SCUBAPRO (Seaglider & Sea Shuttle) and DACOR underwater scooters. Global position satellite (GPS) coordinates from previous studies conducted in 1990 (see Friedlander et al 1994) and in 1996 (see Friedlander 1997) were used to

locate the beginning point of each survey line (Appendix 1). At the start of each site a diver was deployed with a line attached to a buoy. The buoy was used for tracking the divers, and indicating the starting and ending point of each transect so that GPS coordinates could be recorded (from the boat). A Wide Area Augmentation System (WAAS) Garmin GPS (GPSMAP 76) was used to increase accuracy from ± 30 m with a standard GPS unit to ± 3 m (GARMIN Corporation 2001).

Two divers performed each transect. One diver drove an underwater scooter and followed a fixed compass heading recorded in previous USVI transect studies (Friedlander 1997) for a period of at least 30 minutes depending on scooter battery life, underwater conditions, and conch abundance (time needed to measure each conch). This diver also towed a buoy on a line so that the support vessel could track the movement of both divers. A second diver on another scooter was responsible for recording data (see data recording on scooter transect section below).

At the completion of a transect, the diver carrying the marker buoy weighed it down with a dive weight and both divers surfaced. At this point the support vessel recorded the GPS coordinates at the location of the marker buoy indicating the end of the transect line. Transects were conducted during the same time of year (summer months) as those performed in previous surveys (Friedlander 1997, Friedlander et al 1994, Boulon 1987, and Wood and Olsen 1983) to reduce seasonal variability between surveys. Dives did not exceed the no-decompression limits.

Data recording on scooter transects

Transect time was typically about 30 minutes. During each survey, the scooter was kept approximately one meter above the substrate. Divers focused on a width of 2 meters to either side of the scooter path (4 meters). The total distance covered by scooters ranged from 0.1 to 1.5 km. The differences in distances are attributed to the time required to stop and measure conch, as well as suitability of the dive conditions. Underwater visibility, current, and scooter battery life were factors that limited survey time. In areas with no conch, divers did not stop and scooters were operated continuously. In areas with abundant conch, divers stopped while each conch was measured.

As stated above, the first diver maintained compass bearings with the lead scooter. The second diver recorded the transect start and stop times and depths, depth and time of each habitat change, habitat type, and length and lip thickness of each conch were measured and recorded. The first diver also assisted in spotting and gathering the conch for measurements.

Conch length measurements were taken from the apex of the spire to the end of the siphonal notch. Measuring procedures are based on Appeldoorn (1994b). The length of the individual conch was measured to the nearest cm (Friedlander 1997) with an underwater ruler. Lip thickness of adult conchs was measured in the mid-lateral region of the shell with a caliper to the nearest mm. Based on these lip measurements, conch maturity could be determined (see below).

The habitats observed were categorized into six major habitats: (1) unconsolidated sediment (sand or mud with <10% submerged vegetation); (2) coral reef; (3) colonized pavement; (4) rubble; (5) seagrass (continuous seagrass: 90% to 100% cover; patchy seagrass: 10% to <90% cover); (6) algal plain (continuous macroalgae: 90%-100% cover; patchy macroalgae 10% to <90% cover). This benthic habitat classification is based on NOAA's National Ocean Service (NOS) habitat classification (NOS 2000, Appendix 2).

Conch maturity categories

Conchs were separated into five maturity categories based on Appeldoorn (1992, see also Appendix 3):

- Stage 0 (juveniles) any conch without flared lip;
- Stage 1 (newly mature adult) lip thickness 4 to 7 mm, shell clean;
- Stage 2 (sexually mature adult) flared lip fully formed, little to no erosion on Shell, periostracum tan, lip thickness 8 to 15 mm;
- Stage 3 (old adult) outer lip starting to erode, periostracum starting to erode, lip thickness 16 to 33 mm;
- Stage 4 (very old adult = roller) lip very thick and square, heavy erosion present, lip dark rose to platinum gray in color, lip thickness 34 to 59 mm.

2001 Data analysis

The length of each transect was obtained by calculating the distance between the WAAS GPS coordinates of the beginning and end points of each transect. By using the length of transect and time taken to perform each transect, the average speed was calculated in meters/minute. Total area for each transect was calculated by multiplying the length of each transect by the transect width (4 meters). Total areas surveyed are presented in Appendix 4.

Five depth ranges were selected: (a) 0 to 6 m, (b) 7 to 12 m, (c) 13 to 18 m, (d) 19 to 24 m, and (e) 25 to 30 m. Depths were rounded to the nearest whole number. For example, if a depth was recorded at 6.4, then it would be in depth range (a).

If the entire transect depth was within a single depth range (as specified above), then the entire transect length was assigned to that single depth range. However, when the depth in a single transect line changed between different depth ranges (as specified above), then the time duration between recording events (for habitat change or conch occurrence) was divided by the number of different depth ranges that had been encountered (for two depth ranges, time duration was divided by two (=number of depth ranges encountered)). The quotient is then multiplied by speed to estimate the length of the depth range. The area of a depth range is then calculated by multiplying the length of the depth range by the width (4 m).

Calculating areas for each habitat type were simpler than those for the five depth ranges (see above paragraph). When the habitat in a transect line changed, time was recorded. For each transect habitat, the total time in a habitat was calculated (end time minus start time in a single habitat). Total time in a habitat was then multiplied by the estimated scooter speed to get the

total transect length within a single habitat. Once lengths of each habitat were calculated, then area for habitat could be calculated by multiplying the length times width (4 m).

Overall conch densities were calculated by summing the number of conch observed per transect divided by the total area of each transect. Size class densities (adult and juvenile conch) were also determined. Conch densities for each habitat were derived by dividing the number of conch per habitat type by the total area of habitat type per transect. Densities based on depth were calculated by determining the number of conch in each depth range divided by the total area for each depth range in square meters.

Length (juvenile and adult) and age (adult) frequencies were determined. Differences in density or abundance by habitat, depth, or location were analyzed.

RESULTS

2001 Conch densities on scooter transects (including Saba Island, St. Thomas)

A total of 244 queen conch were observed on scooter transects around the USVI conducted in 2001 (Table 3)**.** Nearly 63% of all conch encountered were found around the island of St. Croix although more area was sampled around St. Thomas and St. John combined.

Average densities of adult conch off St. Thomas (24.0 adult conch/ha) were noticeably higher than from St. John (7.2 adult conch/ha). Densities of adult conch varied from 115.9 adult conchs/ha around Water Island, St. Thomas to 0.0 adult conchs/ha found on 14 transects around both St. Thomas and St. John (Figures 2 and 3). Two Brothers, St. John (73.3 adult conch/ha) along with Lindquist Beach, St. Thomas (51.6 adult conch/ha) and West End (57.9 adult conch/ha) had high numbers of adult conch relative to the other survey sites. Saba Island, St. Thomas (a new site added in 2001) had both the highest juvenile conch per hectare and total conch per hectare (284.1 conch/ha) among all survey sites for St. Thomas and St. John in 2001 (Figures 2 and 3).

The densities of adult and juvenile conch in St. Croix differed greatly from St. Thomas and St. John (Table 4). Densities of adult conch in St. Croix varied from 127.7 adult conchs/ha at site 9 to 0.0 adult conch/ha found on 7 of the 16 transects (Figure 4). In St. Croix, the average density of juvenile conch (72.3 juvenile conch/ha) was much greater than that of adult conch (27.4 adult conch/ha). Juvenile densities ranged from 487.0 juvenile conchs/ha to 0.0 juvenile conchs/ha found on 8 of the 16 sites surveyed.

St. Thomas' average conch densities for all transects including those transects common among years and the Saba Island transect were 24.0 adult conch/ha and 27.5 juvenile conch/ha. St. Johns' average conch densities were 7.2 adult conch/ha and 7.5 juvenile conch/ha. On St. Croix, juvenile conch density (72.3 juvenile conch/ha) was nearly three times the density of adults (27.4 adult conch/ha, Figure 5).

2001 Comparisons of densities inside and outside VINP

Of the 12 scooter transects conducted around the island of St. John, six transects (Figure 1 and Table 2) were within the Virgin Islands National Park (VINP). The mean density of adult conch on these transects was 0.0 adult conch/ha (VINP) compared to 14.5 adult conch/ha for the remaining transects around St. John (outside VINP). The mean density of juvenile conch respectively was 10.5 juvenile conch/ha (VINP) and 4.5 juvenile conch/ha (outside VINP). Both sets of transects had high variances due to many transects with zero conchs. The coefficient of variation (CV=SD/mean) for VINP transects was 1.99 while the other locations around St. John had a slightly lower CV of 1.64. Of the 10 juvenile conch observed in VINP, 9 of them (90%) were found on the single transect in Lameshur Bay. No adult conchs were found in VINP. Two Brothers (6 adults) followed by Turner Bay (1 adult) had the only adults observed on all transects.

Comparison of densities for common transects around St. Thomas and St. John between 1996 & 2001 surveys (not including Saba Island, St. Thomas)

In 1996, for common transects around St. Thomas, mean density of adult conch (32.2 adult conch/ha) and juvenile conch (31.5 juvenile conch/ha) was greater (Friedlander 1997) than 2001 (24.2 adult conch/ha; 1.9 juvenile conch/ha). This difference in density between 1996 and 2001 is due almost exclusively to the increased number of transects with zero conchs in 2001 compared with 1996. In 1996 around St. John, density of adult conch was greater (14.7 adult conch/ha) than 2001 (7.2 adult conch/ha), while density of juvenile conch in 1996 (4.1 juvenile conch/ha) was less than that of 2001 (7.5 juvenile conch/ha) (Tables 3 and 4, see Friedlander 1997).

The average density of conch for St. Thomas and St. John was lower in 2001 (2001=19.9 conch/ha; 1996=39.2 conch/ha (Friedlander 1997). Again, the difference in density is due to the absence of conchs observed on most transects in 2001. In 1996 there were fewer transects with 0.0 conch/ha (Tables 3).

Comparison of densities for St. Thomas and St. John among all survey years

In St. Thomas, USVI, data on adult conch densities were only available for 1990 (Friedlander et al 1994) 1996 (Friedlander 1997), and 2001. In St. John, adult conch densities were compared for the nine transects common to all the survey years: 1981, 1985, 1990, 1996 and 2001. Juvenile data was not recorded for St. John in 1981 (Wood and Olsen 1983) and 1985 (Boulon 1987). Mean densities of adult queen conch in St. John were lower in 2001 compared to all other survey years: 1996 (Friedlander 1997), 1990 (Friedlander et al 1994), 1985 (Boulon 1987), and 1981 (Wood and Olsen 1983) (using only the nine transects common to all survey years, see Figure 6). These results are different from the previous comparisons among survey years because the Lameshur Bay, Two Brothers, and Mingo/Lovango transects are excluded from the analysis since they were not surveyed in 1985.

In a previous study, the high variance associated with conch surveys resulted in no statistical difference between surveys conducted in 1981, 1985, 1990, and 1996 using standard statistical techniques (Friedlander et al 1994). However, a Jonckheere test showed a continuous decline in abundance was significant (Friedlander 1997). Although not statistically tested here, there was a decrease in conch density from 82.7 conch/ha in 1996 to 40.7 conch/ha in 2001 for St. Thomas and St. John. However, density of conch in 2001 was one and a half times higher than conch density in 1990 (27.4 conch/ha). This fluctuation reveals the irregularity in conch stocks over time and the need for continued conch stock monitoring.

Conch density among the nine transects common to all survey years on St. John is still in decline or has not increased (Figure 6). Three additional transects were surveyed only in 1990 (Friedlander et al 1994), 1996 (Friedlander 1997), and 2001 (this study). Among these years, 1996 (Friedlander 1997) had the highest adult conch density with 14.7 adult conch/ha, followed by 1990 (Friedlander et al 1994) with 12.6 adult conch/ha, and 2001 with 7.2 adult conch/ha. However, in 2001 juvenile conch density in St. John was greatest (7.5 juvenile conch/ha), followed by 1996 (Friedlander 1997) with 4.1 juvenile conch/ha, and 1990 (Friedlander et al 1994) with 1.4 juvenile conch/ha (Figure 7).

Mean densities among survey years (1990, 1996, and 2001) around St. Thomas varied. In 1990 (Friedlander et al 1994), there were 11.8 adult conch/ha and 1.6 juvenile conch/ha. In 1996 (Friedlander 1997), the mean densities of adult conch were approximately three times higher (32.2 adult conch/ha) and juveniles were approximately 20 times higher (31.5 juvenile conch/ha). In 2001, there was a decline in mean adult density from 1996 to 24.2 adult conch/ha. In 2001 juvenile density in St. Thomas declined to nearly the 1990 level with a density of 1.9 juvenile conch/ha (Figure 8).

1981 vs. 2001 surveys for St. Croix

Surveys of conch abundance around St. Croix, USVI were only done in 1981 and 2001. In 1981 (Wood and Olsen 1983), conch data for all transects were lumped (only densities by habitat type were given). Therefore, there was no data specific to any of the 22 transects. Also, densities of only adult queen conch were provided. Furthermore, only transects 1 to 16 of the original 22 transects were resurveyed in 2001. Transects 17-22 were not surveyed. Average weighted density of adult conch in St. Croix in 1981 (Wood and Olsen 1983) was 7.6 adult/ha. This was dramatically lower than the average adult density in 2001 with 27.4 adult/ha (Figure 9).

In 2001, however, overall density (adults and juveniles) in St. Croix was 99.6 conch/ha (Figure 10). This high density is primarily due to a number of transects having very high densities of conch. For example 4 of the 16 transects had overall densities ranging from 195.5 to 502.7 conch/ha: sites 13 (502.7 conch/ha), 2 (296.6 conch/ha), 4 (224.4 conch/ha), 9 (212.9 conch/ha), and 16 (195.5 conch/ha). Site 13 had the highest juvenile density (487.0 juvenile conch/ha) among all transects (Figure 11).

Habitat types and preferences for St. Thomas and St. John

Algal plain and sand were the most abundant habitat types encountered around St. Thomas and St. John on scooter surveys in 1990 (Friedlander et al 1994), 1996 (Friedlander 1997), and 2001. In 2001, for 21 common transects around St. Thomas and St. John, algae constituted 36% of the total habitat and sand constituted 25% (Figure 12, see also Appendix 5). This was followed by seagrass (18%), coral reef (11%), rubble (9%), and pavement (1%). Hurricane Marilyn in 1995 did not appear to alter the habitat areas surveyed between 1990 and 1996 (Friedlander 1997).

There were noticeable differences in conch densities among habitat types. In 2001 for St. Thomas, seagrass had the highest density of queen conch, followed be algae and rubble (Figure 13). A similar pattern occurred in 1996 (Friedlander 1997). In 1990 (Friedlander et al. 1994), however, rubble had the highest density of queen conch, followed by seagrass, algae, sand, coral reef, and pavement In 2001, pavement was only recorded at one site on St. Thomas and nine sites on St. Croix (Appendix 5). In 2001 for St. John, queen conch was only recorded in seagrass, algae, and sand habitats (Figure 14, and Appendices 6 and 7). Queen conch densities at each site and each island are presented in Appendix 8.

Habitat types and preferences for St. Croix

Pavement and algae were the most abundant habitat types encountered around St. Croix on scooter transects in 2001, constituting 48% and 27%, respectively. All other major habitat types were similar in area surveyed (Figure 15). In 1981, the percentages of habitat surveyed were: algae (41%), followed by coral reef/pavement (26%), sand (18%), and seagrass (15%) (Figure 15, see Wood and Olsen 1983).

In 2001, algal habitats in St. Croix had the highest mean density of conch (113.5 conch/ha), followed by seagrass (94.0 conch/ha) then pavement (25.3 conch/ha) (Figures 16 and 17). When taking the error bars into consideration, there was only a small difference in conch density by habitat type for St. Croix (Wood and Olsen 1983, see also Figure 17).

Depth distribution for St. Thomas and St. John

In 2001, adult conch densities for the original 22 transects around St. Thomas and St. John were highest in the 13 to 18 m depth range (59.0 adult conch/ha, Figure 18). The lowest adult densities were found in the 7 to 12 m depth range (7.8 adult conch/ha) and in the 19 to 24 m depth range (16.7 adult conch/ha). This density distribution was very different for the depth distribution found on transects during the 1996 and 1990 surveys (Figures 19 and 20).

In 2001, juvenile conch showed a decrease in density with increasing depth, while juvenile densities increased with increasing depth for 1996 (Friedlander 1997, Figures 18 and 19). The highest juvenile densities in 2001 were found in the 7 to 12 m depth range (11.1 juvenile conch/ha) while the lowest densities were observed in the 13 to 18 m depth range (5.0 juvenile conch/ha). In 2001, no juvenile conchs were found in the 0 to 6 m, 19 to 24 m, and 25 to 30 m depth range (Figure 18). Contrary to these results, in 1996 the highest densities of juveniles were found at the deepest depth range (31.5 juvenile conch/ha, Figure 19, see Friedlander 1997). Average adult and juvenile queen conch densities by depth for St. Thomas and St. John in 2001 (only) can be seen in Figures 21 and 22 and Table 5. The total area surveyed at each depth range for St. Thomas and St. John is shown in Tables 6.

In 1990 (Friedlander et al 1994), total conch densities were lower at shallow depths (0 to 12 m) and higher at deeper depths (13 to 30 m, Figure 20). Total conch density by depth for common transects around St. Thomas and St. John in 1990, 1996, and 2001 is shown in Table 5.

Depth distribution for St. Croix

No depth data were available for St. Croix in 1981 (Wood and Olsen 1983). However, 2001 data shows that the highest mean adult densities were found in the 19 to 24 m depth range (28.9 mean adult conch/ha), while the lowest were found 7 to 12 m depth (12.5 mean adult conch/ha, Figure 23). There were no queen conch observed in the 0 to 6 m and 25 to 30 m depth range mainly because there were few transects that were completed at these depth ranges. Total area surveyed in each depth range for St. Croix is shown in Table 6.

Size and maturity categories for St. Thomas and St. John

The size distribution of queen conch on 2001 transects showed a bimodal distribution with one peak at 8 cm and a second at 20 to 23 cm (Figure 24). The 8 cm peak consists exclusively of juveniles. The 20 to 23 cm peak consists primarily of conch with a shell lip that is flared. The shell lip typically flares at $3\frac{1}{2}$ years of age (Stoner and Ray 1996) and is a sign of the onset of sexual maturity. Typically conchs that are 3 ½ years of age have a shell length of 20 cm (Creswell and Davis 1991). There were at least 31 mature queen conchs of the 72 observed around St. Thomas (Figure 24) and 7 mature queen conchs of the 19 observed around St. John (Figure 25).

Individual conchs were placed into maturity categories based on shell characteristics (Appendix 3). In 2001, approximately 42% of all conch observed on scooter transects around St. John were juveniles (stage 0). The majority of the adults (40%) were stage 3 individuals that were older adults with the outer lip starting to erode, the periostracum starting to erode, and fouling of the shell occurring. A small percentage (11%) of very old (stage 4) adults were observed in St. Thomas. An even smaller percentage (7%) of newly mature adults (stage 1) was observed. Studies suggest the conch soon grow out of this category (Friedlander 1997).

In St. John, approximately 63% of all conch observed on scooter transects were juvenile (Figure 26). Most adult conchs were stage two (sexually mature) individuals. A small percentage (5%) of conchs was newly mature adults (stage 1). Also, there were no old (stage 3), nor very old (stage 4) adults observed (Figure 26). Shell length frequencies for St. Thomas/St. John in 1996 and 2001 are shown in Figure 27.

2001 size and maturity categories for St. Croix

Queen conch in St. Croix revealed a fairly uniform size distribution (see Figure 28). There were similar proportions of conch at all sizes. Forty-nine conchs of 153 observed around St. Croix were sexually mature (32%). Approximately 68% of all conchs observed on scooter transects were juveniles (stage 0). Among the adult conch, 25% were in stage one, 50% were stage two, 25% were stage 3, and 0% were stage four (Figure 29). Shell length frequencies for St. Croix in 1981 and 2001 are shown in Figure 30.

Shell length frequencies for the territory (St. Thomas, St. John and St. Croix) in 2001 are shown in Figure 31.

DISCUSSION

Comparison of queen conch densities among islands in the US Virgin Islands

During 2001, queen conch densities were higher on scooter transects around St. Thomas compared to St. John. The densities of queen conch observed at each survey site in St. Thomas and St. John in 1990, 1996 and 2001 are shown in Figures 32 to 50. Conch densities, particularly juvenile densities, were higher on scooter transects around St. Croix compared to St. Thomas and St. John (Figure 5). For St. Croix, mean density of adult queen conch increased from 7.6 adult conch/ha in 1981 (Wood and Olsen 1983) to 27.4 adult conch/ha in 2001 (juvenile conch were not sampled in 1981). Around St. Thomas, only two sites had relatively high densities of conch in 2001, Inner Water Island and Saba Island (the new site). Overall, St. Thomas had an average of 24.0 adult conch/ha and 27.5 juvenile conch/ha in 2001. In St. John in 2001, all sites had low conch abundances, and an overall mean density of 7.2 adult conch/ha and 7.5 juvenile conch/ha. Around St. Croix in 2001, several sites had high conch densities, particularly for juveniles. For example, site 2 had 247.0 juvenile conch/ha, site 4 had 138.1 juvenile conch/ha, site 13 had 487.0 juvenile conch/ha, and site 16 had 138.0 juvenile conch/ha. In 2001 overall, St. Croix had 27.4 adult conch/ha and 72.3 juvenile conch/ha. In 1998, 50% of all conchs harvested in St. Croix were juveniles (CFMC 2001). This not only suggests that juveniles in St. Croix are common, but recruitment is strong. However, it also indicates that fishing pressure is high and that stocks may decline because of reduction in sexually mature adults owing to over harvest of juveniles.

The higher conch densities on St. Croix compared to St. Thomas/St. John are likely due to the shallower water depths of the St. Croix shelf and perhaps exploitation patterns. In St. Croix, the shelf area does not extend far from the shoreline $(3 nm), and the shelf depth ranges from 6 to 12$ m. This depth range is primarily where juvenile and subadult size/age categories for conchs are found (Schweizer and Posada, draft).

The shelf around St. Thomas and St. John extends much further (8 nm to the south, and 20 nm to the north), and is primarily between 25 to 45 m deep. Many of the surveys around St. Thomas and St. John were not conducted in deep waters (>24 m) as the survey sites' locations were based on previous studies and dive limitations. Therefore, adult breeding stocks typically found at deeper depths (Appeldoorn 1994a) were not surveyed or encountered. Further resource studies at water depths >24m are needed.

There are restrictions and regulations in effect for the harvest of conch in the USVI. A bag limit of 2 conchs per person per day (recreational fishers) has been in effect in the VINP in St. John since 1962. No commercial take of conch is permitted in the VINP. Despite this regulation, no difference in conch densities were detected between transects conducted in the VINP and other sites outside VINP, St. John.

In St. Thomas and St. John, there was a moratorium on the harvest of queen conch from 1988 to 1992 (see Friedlander 1997). The density of queen conch increased from 13.4 conch/ha in 1990 to 63.7 conch/ha in 1996 on St. Thomas (Table 4). This suggests that the moratorium was effective. However, this does not account for the decline in the density of queen conch on St. John. This decline on St. John may be because about half of the sites were in the National Park which has its own regulations that do not appear to be effectively enforced, and because the moratorium was primarily done at the request of St. Thomas fishers who were therefore more likely to adhere to the moratorium. Few St. John fishers are licensed or participate in the St. John/St. Thomas Fisheries Advisory Committee (FAC).

Current queen conch regulations require fishers to harvest only queen conch that are at least 9" in length or have a shell thickness of 3/8". Conch must be brought to shore whole in the shell. Recreational harvesters can harvest only 6 conch per person or 24 per boat in territorial waters. Commercial fishers can harvest 150 queen conch per commercial fisher per day. These regulations are not effectively enforced because of the lack of enforcement officers dedicated to enforcing marine fisheries laws.

Queen conch recruitment

During the 2001 survey, only 4 juvenile conchs were observed on the scooter transects previously surveyed in 1990 and 1996 around St. Thomas. At the new St. Thomas site, Saba Island, an additional 26 juvenile conch were observed in a shallow water seagrass bed (6 to 10 m). Juvenile and adult conch abundances varied between sites. Given the high degree of variability and reports of high densities of juvenile queen conch in some un-sampled bays (e.g. Brewers Bay, St. Thomas and an unsampled portion of Fish Bay, St. John), it is recommended that funding for this project be increased to allow more adequate sampling of habitats in the Virgin Islands. However the objective of this survey was to detect historic trends at established stations over time, not to identify and survey new areas for conch.

The low numbers of juvenile queen conch found at most sites on St. Thomas and St. John may be owing to lack of successful recruitment. According to Stoner and Ray (1996), at least 53.0 adult conch/ha are needed for optimal reproduction to take place. Therefore many of the sites in St. Croix clearly have sustainable densities. However, most sites around St. Thomas and St. John do not have conch densities high enough for optimal reproduction.

Conch larvae live in the upper water column for 16-28 days, feeding on phytoplankton (Stoner and Ray 1996). Given the range of time for planktonic development before settlement, it is likely that a portion of queen conch larvae returns to their natal habitat and a portion disperse long distances. Genetic studies and analysis of queen conch larvae life history and ocean currents in the Caribbean region suggest that larvae can travel long distances with the currents before settling out of the plankton and metamorphosing (Stoner and Ray 1996). The prevailing water current is from east to west in St. Thomas and St. John. This suggests that recruitment in St. Thomas and St. John may be linked to the reproductive success of queen conch of British Virgin Islands' (B.V.I located east of St. Thomas/St. John) conch stocks. Recruitment in the USVI may be a function of how well the queen conch fishery in the B.V.I is managed and protected.

The low numbers of juveniles in St. Thomas/St. John may also be a function of the illegal harvest of juveniles in the easily accessible shallow coastal waters around the islands. The Division of Fish and Wildlife's knowledge of large populations of juveniles in Brewer's Bay and Fish Bay is a function of telephone calls from members of the public who reported the illegal harvest of juvenile queen conch from these bays. It was estimated that several hundred juvenile queen conch were removed from Fish Bay and the Division of Environmental Enforcement reported over 500 juveniles conchs removed and killed from Brewer's Bay in one incident.

Queen conch abundance related to habitat

Queen conchs were most abundant on algal plains and seagrass beds (Table 7 and Appendix 7). Around St. John, much of the habitat surveyed was shallower than St. Thomas (Appendix 1). Overall in St. John, conchs were found almost exclusively in seagrass beds (63% of all conchs on scooter transects). Such was the case for 1996 surveys (Friedlander 1997) as well. Conchs were found primarily in manatee grass (*Syringodium filiforme*) and turtle grass (*Thalassia testudinum*).

From 1990 (Friedlander et al 1994) to 2001, the percentage of surveyed habitat covered by seagrass has increased, or remained relatively stable among the 22 common transects around St. Thomas and St. John (Figure 12). In fact, the percentage of each of the major habitat types surveyed on St. Thomas and St. John has remained rather uniform from 1990 (Friedlander et al 1994) to 2001 (Figure 12).

In St. Croix habitat was much more patchy in 2001 compared to 1981. Also, new habitats were surveyed in 2001 compared with 1981 (Wood and Olsen 1983, see Figure 15). The habitat patchiness and new habitats may be primarily due to hurricanes and storms within this 20-year period (especially hurricanes Hugo in 1989, Marilyn in 1995 and Lenny in 1999). Other possible contributing factors include anchor damage, water pollution, increased runoff and sedimentation, and global warming. Data on conch density and size/age distribution by habitat are useful for establishing refuges for various life stages of these organisms and estimating densities over the entire shelf area.

The effect of hurricanes on queen conch density

In September 1989, Hurricane Hugo struck the Virgin Islands with wind speeds up to 150 per hour (Quinn 1991). It has been suggested that the effects of the hurricane retarded the conch population recovery. For example, in Lameshur Bay, St. John, the population of conch after the hurricane were 10% of what they were 3 years prior (Quinn 1991). In August 16 1995, Hurricane Marilyn also struck the Virgin Islands with sustained winds of 104 mph with maximum gust to 129 mph (NOAA 2000). Despite Marilyn's devastation, density of conch was greater in 1996 than 1990 for St. Thomas (Table 3). The density increase was possibly a function of the moratorium that was in effect for 5 years.

Age structure of queen conch

Many of the adult conchs observed around St. Thomas were old individuals. Juvenile abundance was extremely low in most locations around St. Thomas (not including Saba Island) and much lower than adult abundance.

Conclusion

In summary, abundance declined between 1996 (Friedlander 1997) and 2001 for common transects in St. Thomas (Figure 8), and between 1981 (Wood and Olsen 1983) and 2001 for the 9 transects common to all survey years for St. John (Figure 6). If management regulations are not enforced, continued decreases in conch abundance are expected in St. Thomas and St. John. St. Croix, on the other hand, has abundant populations of conch at many sites. If laws are enforced and the conch fishery properly managed, continued increases in conch abundance may be achieved.

Any effective management strategy should include prohibition of recreational conch harvest within park boundaries. Reserves have the benefit of providing migrants to surrounding areas, enhancing reproduction, exporting larvae, and providing biological reference areas (Bohnsack 1996). Reserve areas for queen conch in the Bahamas have shown to have 31 times more adult conch than fished areas (Stoner and Ray 1996). Protected sites within St. John's National Park did not demonstrate increased conch abundance. This may indicate that they do not have much protection. This is particularly confounding, because most of the sites included coral sand and seagrass beds, which are used during summer spawning periods (Brownell and Stevely 1981). Additionally, there are large rhodolith/rubble communities surrounding the island of St. John, which provides deeper rubble habitats that are the primary winter feeding habitats for many older conchs (Stoner and Sandt 1992).

Long term restocking of small conch into the natural environment probably is not an economic way to replenish and support the fishery (Appeldoorn 1994a). Restocking the ocean with hatchery-reared conch is a very costly project, considering the high mortality rate within the first 6-9 months of release (Appeldoorn 1994a). The cause of the initial decline in the conch populations is not from natural recruitment failure, but from overfishing and/or destruction of

habitat, which result in recruitment failure (Appeldoorn 1994a). Conch populations can ultimately be restored by effective resource management and heightened enforcement.

All management measures are only as good as the level of compliance. Generally, there are insufficient resources and/or personnel for enforcement purposes (Uwate et al 2001). Enforcement of conch regulations is often viewed in isolation. Allocation of resources toward enforcement should be viewed in terms of maintaining the viability of all marine resources. The worth of all resources is sufficient to warrant an investment in effective enforcement.

Finally, public education is probably one of the most important and long-lasting strategies for managing fishery resources. Community awareness and public support for management strategies can encourage restoration efforts. Cooperative management by everyone that uses conch resources will ultimately benefit all involved groups.

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*notes:

- 1. These maps are not to scale.
- 2. For St. Croix, only transects 1 to 16 were surveyed in 2001.

Note: Saba Islands is a new site added to the St. Thomas transect list in 2001

*Note: Only the first 16 of the original 22 transects were surveyed in 2001

data from Friedlander (1997)

*Note: 1990 and 1996 data from Friedlander (1997)

*Notes: 1) 1990 and 1996 data from Friedlander (1997)

2) No available data for St. Thomas in 1981 and 1985

3) Error bars are standard error of the mean

*Notes: 1) 1981 data from Wood and Olsen (1983)

2) 1981 data provide an average weighted density

3) No standard deviation provided for 1981

*Notes: 1) 1981 data from Wood and Olsen (1983)

2) Error bars are standard deviation of the mean

3) 1981 data provide an average weighted density

4) No standard deviation provided for 1981

*Notes: 1) 1990 and 1996 data from Friedlander (1997) and Friedlander et al (1994) 2) Sand (SD), Coral Reef (CR), Rubble (RB), Seagrass (SG), Pavement (PV), Algae (AL)

*Note: Sand (SD), Coral Reef (CR), Rubble (RB), Seagrass (SG), Pavement (PV), Algae (AL)

- 2) In 2001 only 16 of the original 22 transects were surveyed
- 3) Sand (SD), Coral Reef (CR), Rubble (RB), Seagrass (SG), Algae (AL)
- 4) Coral Reef habitat includes Pavement (PV)

*Note: Sand (SD), Coral Reef (CR), Rubble (RB), Seagrass (SG), Pavement (PV), Algae (AL)

- 2) Sand (SD), Coral Reef (CR), Rubble (RB), Seagrass (SG), Algae (AL)
- 3) Coral Reef habitat includes Pavement (PV)

*Note: No depth data available for 1981 and 1985

*Note: Sexual maturity reference from Stoner and Ray (1996)

*Note: Sexual maturity reference from Stoner and Ray (1996)

*Note: Maturity categories based on Friedlander (1997)

*Notes: 1) 1996 data from Friedlander (1997) 2) Sexual maturity reference from Stoner and Ray (1996)

*Note: Maturity categories based on Friedlander (1997)

- 2) 1981 data from Wood and Olsen (1983)
- 3) 1981 shell length size-frequency distributions were provided for juvenile conch only

*Note: Sexual maturity reference from Stoner and Ray (1996)

*Notes: 1) 1990 data from Friedlander et al (1994) 2) 1996 data from Friedlander (1997)

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*Notes: 1) 1990 data from Friedlander et al (1994)

2) 1996 data from Friedlander (1997)

3) No juvenile data available for 1981 and 1985

*Notes: 1) 1990 data from Friedlander et al (1994)

2) 1996 data from Friedlander (1997)

3) No juvenile data available for 1981 and 1985

*Note: All data from State/Federal catch reports submitted to the Division of Fish & Wildlife, DPNR and NMFS. Weights are in pounds of meat.

Note: Asterisks () denote location within Virgin Islands National Park.

*Notes:1) 1996 data from Friedlander (1997), 1990 data from Friedlander et al. (1994).

2) No site data available for St.Thomas in 1981 and 1985, only St. John was surveyed.

3) Saba Island is a new site added in 2001.

 *Notes: 1) 1981 data from Wood and Olsen (1983), 1985 data from Boulon (1987), 1990 and 1996 data from Friedlander (1997). 2) No juvenile data was available for queen conch in 1981 and 1985.

*Notes: 1) No density data available for individual sites in 1981 (total mean density $= 7.6$ conch/ha).

2) No surveys were conducted in 1985, 1990, and 1996.

3) No surveys were conducted in 1985, 1990, and 1996.

4) Totals for density columns are mean values.

*Notes: 1) 1981 data from Wood and Olsen (1983), 1985 data from Boulon (1987), 1990 data from Friedlander et al. (1994), Friedlander (1997).

2) Wood and Olsen (1983) indicate that the average weighted adult density for St. Thomas and St. John is 9.7 adult/ha

3) Friedlander et al. (1994) was the only report that provided raw densities from St. John transects performed in 1981

4) There is no raw data available for St. Thomas in 1981 and 1985.

5) In 1985 only St. John was surveyed.

*Note: 1990 data from Friedlander et al. (1994), 1996 data from Friedlander (1997).

*Notes:

- 1. 1981 data from Wood and Olsen (1983), 1990 data from Friedlander et al (1994), 1996 data from Friedlander (1997);
- 2. Rubble (RB) and Pavement (PV) were categorized as Coral Reef (CR) by Wood and Olsen (1983);
- 3. There is no habitat data available for 1985.

*Note: 1981 data from Wood and Olsen (1983).

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Note: Habitat classification based on NOAA's National Ocean Service (NOS) habitat classification (see NOS 2000).

*Note: Maturity categories from Friedlander (1997).

*Notes: 1) 1981 data from Wood and Olsen (1983), 1985 data from Boulon (1987), 1990 data from Friedlander et al. (1994), 1996 data from Friedlander (1997).

 2) In 1981 surveys were completed using towed-diver transects. Scooter transects were used for all other survey years.

